

L. J a k u c s

THE KARSTIC CORROSION OF NATURALLY OCCURRING
LIMESTONES IN THE GEOMORPHOLOGY OF OUR AGE

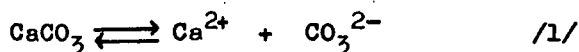
During recent years the theory of karst corrosion has been considerably modernised, partly due to the new interpretations of the processes resulting from the rapid development of chemistry, partly, however, as a consequence of the ever-increasing mass of karst research data. Thus, the zonal and extrazonal richness of form of the karsts genetically can today be explained only on the basis of the accurate knowledge of the nature and extents of the processes involved. Contributions have been made to the development of certain aspects of the theories of dissolution by the results of Hungarian research, including those of the present author. Hence, the review of this subject was considered justified.

The concept of karst corrosion is taken to understand primarily the particular dissolution denudation processes of limestones /and also, in a wider meaning, those of other limestones which dissolve well in water without a residue/.

In connection with the dissolution of limestone in water, three quite different processes must be considered. First, a distinction must be made between carbonate dissolution in pure /distilled/ water, and hydrocarbonate dissolution, that is the dissolution of limestone in water containing carbon dioxide. There are specific qualitative and quantitative differences between the two, but both processes are reversible. The third type of dissolution involves the participation of other chemical agents, mainly organic and inorganic soil acids, etc. In this latter case the new calcium compound formed in solution is generally no longer able to reform CaCO_3 .

I. The carbonate dissolution of limestone

When limestone CaCO_3 dissolves in distilled water, its dissociated ions remain unchanged /apart from solvation/, and the act of dissolution can be expressed by the following reversible reaction:



An equilibrium is established between the solid phase CaCO_3 , the solvent H_2O and the dissolved ions, and this can be defined by the solubility product K_1 /TILMANS 1942, MILLER 1952/. K_1 is a variable index, the absolute value of which generally changes linearly with the temperature of the solution, but it also depends on certain crystal-structural properties of the solid phase. The K_1 for the rhombic dipyramidal modification of CaCO_3 /aragonite/ is higher than that for the trigonal calcite /SCHOELLER 1956/.

Table 1 is a collection of the literature data which give the absolute and variable orders of magnitude for K_1 .

Table 1

Author /year/	t /°C/	dissolved CaCO ₃ in mg/l		
		calcite	aragonite	amorphous CaCO ₃
SCHLOESING ^x /1872/	8,7	10,0	-	-
SCHLOESING /1872/	16,0	13,1	-	-
SCHLOESING /1872/	25,0	14,3	-	-
CHARLOT- ENSCHWILLER /1939/	-	12,4	-	-
TROMBE /1952/	16,0	16,0	-	-
HODGMAN /1955/	cold	14,25	-	-
HODGMAN /1955/	warm	18,75	-	-
SCHOELLER /1956/	25,0	14,33	15,28	14,45
SCHOELLER /1956/	50,0	15,04	16,16	15,15
SCHOELLER /1956/	100,0	17,79	19,02	18,16

^x The pedological work of SCHLOESING, which appeared in 1872, is not known in the original in Hungary; nevertheless its data are frequently cited by Western authors, e.g. O. LEHMANN /1932/, TROMBE /1952/, etc. The data of SCHLOESING in the Table are therefore given only on the basis of secondary sources.

The views to be found in the literature regarding the time-requirements for the dissolution of limestone in pure water are fairly contradictory. According to HARRASOWITZ 1954, BÖGLI 1956, 1960, H. LEHMANN 1956, 1960, BAUER 1964, FRANKE 1967, etc., the reaction concerned must be extremely fast. This conclusion was based on the fact that when the above authors analyzed samples of water collected from limestone surfaces during field-studies they were unable to detect dissolved CaCO_3 in concentrations lower than those given in the Table. BAUER /1964/, for example, collected for analysis the moisture trickling down "in statu nascendi" in a karr-channel; this had been in contact with the surface of the limestone for only a few seconds. Accordingly, BÖGLI /1960, 1963/ denotes the time-requirement for non-carbonate primary dissolution as about 1 second. /See also later in connection with the hydrocarbonate dissolution./

In the view of MARKÓ /1963/, however, these experimental results confirm not the high rate of dissolution, but the fact that the ion-diffusion under the special experimental conditions was particularly rapid; nevertheless, the rate of this process, like the dissolution of all solid substances, is controlled primarily by the rate of diffusion of the dissolved molecules or ions from the saturated solution layer in the vicinity of the surface of the solid phase to the more distant solution layers. It is natural, therefore, that the diffusion conditions will be optimal in a thin liquid film flowing over the surface; in accordance with the MARKÓ interpretation, this can explain the greater intensity of the dissolution dynamics.

At the same time, however, GERSTENHAUER and PFEFFER /1966/ go still further and, on the basis of their own extensive investigations, directly cast doubt on the reliability of BÖGLI's analytical data. They had collected limestone samples from various places throughout the world, and checked their time-curves under sterile laboratory conditions; they did not observe such a rapid achievement of saturation, but instead /as shown in Figure 1/ times which were longer by several orders of magnitude. In some samples for example, the time during which the sample was in contact with the water, and which was necessary to attain a solubility of 13 mg/l, was a whole day; this is even more surprising, because the water used contained a little carbon dioxide, which normally accelerates the dissolution.

The cause of these sharp, apparently irreconcilable differences here lies almost certainly in the different chemical investigation methods used by the various authors. Without going into details /these are given by MERCK, MÜLLER and GERSTENHAUER/, it must be pointed out that BÖGLI and FRANKE titrated their water samples by the WARTHA-LUNGE method with methyl orange as indicator, whereas GERSTENHAUER and PFEFFER used titriplex III /chalconecarboxylic acid/. The viewpoint of GERSTENHAUER is in all probability closer to the truth, because the indicator used is definitely calcium-specific; it reacts only with Ca^{2+} ions and gives a reliably clear colour change at the end-point even with quite weak solutions.

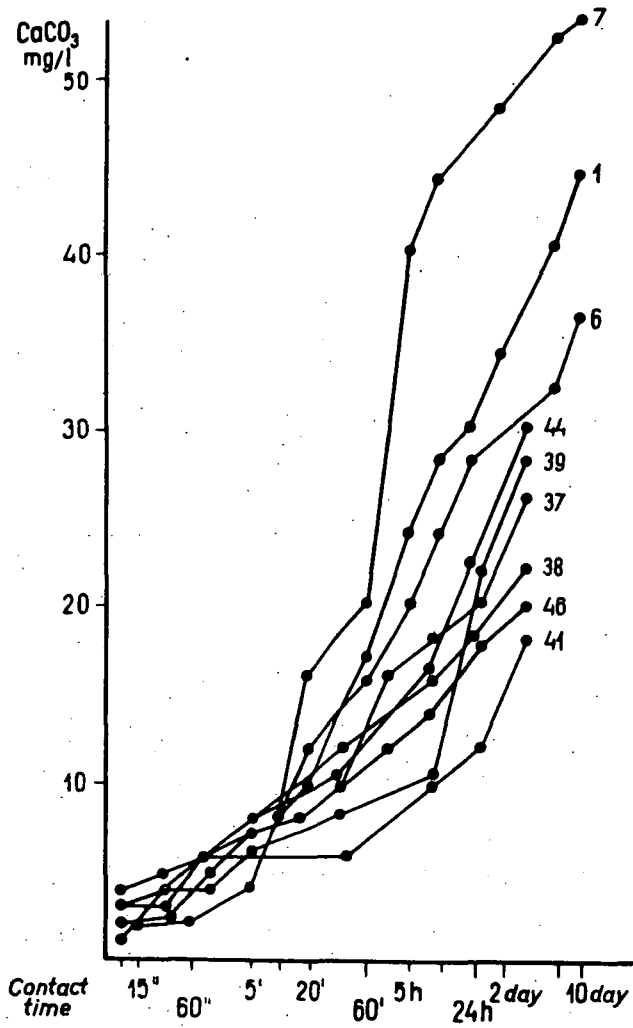


Figure 1. Time courses of the dissolution of various types of limestone in water saturated with atmospheric CO_2 /according to GERSTENHAUER and PFEFFER/.

This carbonate dissolution process is of fairly subordinate practical importance in nature of course, at least directly. Completely pure /quasi-distilled/ water does not exist under the geographical conditions leading to karstification. The freatic waters which must be considered from the point of view of actual karstic corrosion /but rain-water too/ always contain certain amounts of dissolved chemical substances, and the variable amounts and compositions of these change the capacity to dissolve carbonate in a decisive way.

Of these factors, the role of the carbon dioxide saturation of the water is stressed in the literature, since water containing only a little carbon dioxide will dissolve many times the amount of limestone dissolved by the same volume of pure water.

II. The hydrocarbonate dissolution of limestone

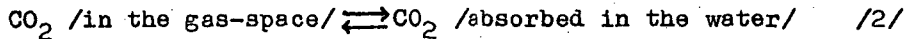
The dissolution of limestone by carbonic acid is termed hydrocarbonate dissolution, because the carbonic acid ions present in the solution react with the limestone to yield the well-soluble calcium hydrocarbonate, $\text{Ca}/\text{HCO}_3/2$ /KYRLE 1923, TILMANS 1932, etc./.

The factor controlling the dissolution of limestone as hydrocarbonate, and the opposite of the dissolution process, the redeposition of the limestone from the solution, is in all cases the carbon dioxide /carbonic acid/ concentration of the water /or solution/ concerned. To a first approximation it can be said that the water in contact with the limestone is capable of dissolving more CaCO_3 , the greater the concentration of the CO_2 previously dissolved in it.

If the CO_2 content of the water in contact with the limestone is increased, therefore, the amount of limestone dissolved will increase, while if the carbon dioxide leaves the solution /e.g. by evaporation/, the dissolved limestone content of the solution will decrease by means of its precipitation from the solution and its accumulation /lime-muds of chemical origin, stalactites, stalagmites, travertine, etc. are formed in this way/.

The total dissolved CO_2 content of some natural water may vary considerably, and since there is a direct connection between the CO_2 content of the water and its ability to dissolve carbonate in order to interpret the corrosion dynamism of the given water it is of the greatest importance to study all those factors which affect the amount of CO_2 dissolved by the water in contact with the limestone.

The following equilibrium is involved:



The amount of carbon dioxide absorbed is affected by the following factors:

1. The carbon dioxide concentration of the gas phase in contact with the water. This is expressed by the partial pressure of the carbon dioxide /p CO_2 /.
2. The mutual temperature of the water and of the gas, at which the process of dissolution occurs.
3. The hydrostatic pressure acting jointly on the water and the gas phase in contact with it.

4. The amount of time available for the dissolution process.

Thus, it can readily be seen that the dissolution of gaseous CO_2 and the decomposition of the solution /reformation/ depend in an extremely complex way on the combination of the above factors, for the change of any one of them can displace the equilibrium in either direction.

Although the conditions controlling the dissolution of CO_2 in nature are so complex, it is nevertheless worthwhile, for the sake of clarity, to consider the above four points separately.

It should first be pointed out that in accordance with the HENRY-DALTON law /1803/ the amount of the gaseous component dissolving in the liquid phase in contact with the gas is directly proportional to p , the partial pressure of that component, and is inversely proportional to the change of temperature. Thus, the amount of CO_2 dissolved in the water can be determined with the help of the following formula:

$$\text{dissolved } \text{CO}_2 \text{ in g/l} = L \cdot p \cdot 1.9634 \quad /3/$$

where the value 1.9634 is the reciprocal density of CO_2 in l/g, p is the partial pressure of the CO_2 , and L is the solubility or absorption coefficient for the CO_2 gas, which varies with the temperature. Table 2 contains the values of L for the more important temperatures involved in the karstification.

Table 2

	dissolution temperature, °C							
	0	5	10	15	17	20	25	30
L =	1,713	1,424	1,194	1,019	0,958	0,878	0,765	0,665

Equation /3/ was used as long ago as 1872 by SCHIOESING, who calculated and tabulated the amounts of carbon dioxide to dissolve in water at given temperatures and p_{CO_2} values. The most relevant of his values are given in Table 3.

It can be seen from Table 3 that at 5 °C, for example, with a typical p_{CO_2} value for the atmosphere of 0.0003 /0.03 % carbon dioxide content/, the moisture in contact with the air is capable of absorbing about 0.84 mg/l carbon dioxide. At higher temperatures, for instance at 30 °C, which occurs in Hungary too during the summer, but which is general in the tropics, the amount of CO_2 absorbed is less /0.39 mg/l/.

Table 3

The change of the CO₂-dissolving capacity of water with the temperature and p CO₂ values occurring in practice

p CO ₂ in the atmosphere	Total CO ₂ dissolved in the water /in mg/l/, at the following temperatures							
	0 °C	5 °C	10 °C	15°C	17°C	20 °C	25° C	30 °C
0,0001	0,34	0,28	0,23	0,20	0,19	0,17	0,15	0,13
0,0003	1,01	0,84	0,70	0,60	0,56	0,52	0,45	0,39
0,0005	1,68	1,40	1,17	1,00	0,94	0,86	0,74	0,65
0,00075	2,52	2,09	1,76	1,50	1,41	1,28	1,12	0,98
0,001	3,36	2,80	2,34	2,00	1,88	1,72	1,49	1,31
0,0015	5,04	4,19	3,51	3,00	2,82	2,58	2,24	1,96
0,002	6,73	5,59	4,69	4,00	3,76	3,45	3,01	2,61
0,0025	8,40	6,99	5,85	5,00	4,70	4,28	3,72	3,28
0,005	16,8	13,98	11,7	10,0	9,4	8,57	7,46	6,56
0,0075	25,2	20,9	17,6	15,0	14,1	12,8	11,2	9,79
0,01	33,6	28,0	23,5	20,0	18,8	17,2	14,9	13,1
0,02	67,3	55,9	46,9	40,0	37,6	34,5	30,0	26,1
0,03	101,	83,9	70,4	60,0	56,6	51,7	45,1	39,2
0,04	135,	112,	93,8	80,1	75,6	69,0	60,0	52,5
0,05	168,	140,	117,	100,	94,1	86,2	74,6	65,3
0,06	202,	168,	141,	120,	113,	103,	90,0	78,4
0,07	236,	196,	164,	140,	132,	121,	105,	91,4
0,08	269,	224,	188,	160,	151,	138,	120,	104,
0,09	303,	252,	211,	180,	169,	155,	135,	118,
0,1	336,	280,	235,	200,	188,	172,	149,	131,
0,2	673,	559,	469,	400,	376,	342,	300,	261,

At the same time, however, it is true that in the tropics, where an intense CO_2 production takes place during the entire year as a result of the decomposition of the rich organic matter and also because of other soil and biogenetic processes, the carbon dioxide content of the air in the vicinity of the soil can attain values somewhat higher than the average. For example, LEHMANN /1955/ determined a value of 2.5 mg/l for the CO_2 content /thought to be of atmospheric origin/ of rain-water collected from air at temperature of 22 °C. This would correspond to a partial pressure of CO_2 in the atmosphere of about 0.0016 according to the SCHLOESING tabulation /Table 3/ /H. LEHMANN 1956/.

The surprisingly large atmospheric carbon dioxide content calculated from this /0.15 %/, however, was not controlled directly by means of a simultaneous air-analysis, and so the relatively really high dissolved CO_2 value can not be sufficiently convincingly documented as being exclusively of an atmospheric diffusive origin. It is not excluded, for example, that the CO_2 content of the water may have been increased by the mixing with it of fine particles of spray arising from the high-energy collision of the rain-drops with the soil, since it is known that the water component of the aerosols so formed contains ions from the soil in relatively high concentrations /CAUER 1954, JAKUCS 1953, 1959/. A further contributory factor may be that the p. CO_2 values of the rain are determined not by the observed temperature of the precipitation zone, but by the much lower temperatures prevailing at the cold vapour-condensation cloud level at a height of several thousand metres.

The fact that the factors mentioned almost certainly do modify the chemical composition of the rain-water collected fairly close to the soil surface, and to a considerable extent /an order of magnitude due mainly to the aerosol factor/, is supported by similar water-analysis

results from the temperate zone, and even from high mountains. Such studies were carried out in the Alps by BÖGLI /1960/ and BAUER /1964/, and in Hungary by CZÁJLIK /1961/. Practically agreeing amounts of carbon dioxide were found volumetrically in rain-water samples collected above the soil /in the range 1.32-3.63 mg/l CO₂/.

Of course, one can hardly talk of a very significant hydrocarbonate dissolution of the limestone even in the presence of 2-3 mg/l dissolved CO₂, for /as can be seen from Table 4/ in order for the water in contact with the limestone to dissolve sufficient of it as hydrocarbonate so as to accord with the simple dissociation given in equation /1/, i.e. about 13 mg/l CaCO₃, a dissolved CO₂ content of about 6 mg/l would be necessary at 17 °C; this is equivalent to approximately ten times the p CO₂ to be found in the atmosphere.

In contrast with this, the actual dissolved CO₂ content of the soil waters and karst waters occurring in nature, and which are responsible for the karstic corrosion, is in fact always many times higher /it is very often more than 100 mg/l even!/. There is no doubt that the rain-water which falls onto the soil obtains this high carbon dioxide concentration not from the air, but from the soil itself.

The fact that soils, and thus the soil species covering limestones too, possess a soil atmosphere with a high p CO₂ value, was demonstrated long ago by pedological research workers /SCHLOESING 1872, FODOR 1875, WOLLNY 1880, BOUSSINGAULT and LÉVY 1853, etc./.

The study of the soil atmosphere, that is the gas mixture occupying the pore volume in soils with looser structures, is dealt with in detail later. At this point, however, it should be mentioned that the pedological investigations have indicated a CO_2 content of more than 1 % in the atmospheres of humous soils rich in organic matter, while it may frequently even be more than 10 %. This must be regarded as the most decisive factor governing the dynamism of the karst corrosion, for the rain-water falling onto the soil surface and absorbed there will be affected immediately by the soil air which comes into contact there on a very large surface with the water adhering to the soil particles, and thus even in the uppermost soil layer the water will become saturated with carbon dioxide in accordance with the $p \text{CO}_2$ value of the gas mixture there.

The very important role of the soil atmosphere in regulating the dissolution of the limestone was also observed relatively early by geographers investigating the genesis of karsts. KNEBEL analyzed the interaction quite comprehensively as long ago as 1906, but the relations between the dissolving power of the water and the soil had been considered long before him. For example, in 1831 the Hungarian VASS had given an already detailed account of the stalactite and stalagmite formation theory of PARROT, LANG and SOMMER, which is a reasonably good approximation to the presently accepted scientific view. According to VASS, these researchers explain the stalactite formation by the dissolving work of the rain-waters seeping through "the layer of decayed plants and vegetable mould which is saturated with carbonic acid".

CHOINOKY /1940/ also stressed the role of the carbon dioxide of the soil in the aggressiveness of the water: "It is known that a constant decay and a slow oxidation take place in the soil, with the formation of carbon dioxide. This is then concentrated on the soil particles, since these always have gas-concentrating natures. If it begins to rain after dry weather, the "smell of rain" can be detected. This is the characteristic smell of those gases which are expelled from the surface of the particles of soil. It can be demonstrated that at such time there is very much gaseous carbon dioxide in the lower layers of the air of the wetted area. Here, therefore, much gaseous carbon dioxide gets into the fallen rain-water, especially as a result of the dynamic pressure of the colliding rain-drops. This considerable amount of gas is carried by the water into the cracks in the rock." /op. cit. p. 1006./

The measurements of TROMBE and of JACKLY in France and Switzerland /TROMBE 1951, 1952, 1956/ indicated that the localization of CO₂ gas in the soil layers covering the karst formations can reach proportions /10-25 %!/ which are never observed otherwise in the atmosphere.

Very extensive and thorough studies were carried out in Hungary by FEHÉR /1954/ in order to discover the variations of the CO₂ content of the soil atmosphere and the regularities involved, while the present author also made measurements on Hungarian and Yugoslav karst soils in 1967-8. The results of all these investigations show that the composition of the soil atmosphere reacts with very rapid and sensitive changes to both the macro- and microclimatic factors, and characteristic simultaneous

differences are shown even within a given study site /e.g. a single sink-hole/, depending on the type of vegetation living in the soil, and even on the individual species in the rhizosphere /between the roots of the plants/ JAKUCS 1970, 1971.

It clearly emerges from what has been said that the permeating waters originating from the rainfall, which direct the entire karst development by means of their dissolution of the limestone, acquire their carbonic acid content /which determines the dynamics of the corrosion/ essentially always and everywhere in the uppermost layer levels of the soil. This means therefore that in a certain region the rate of the corrosion karst-denudation is controlled most importantly by /in addition to the amount of rain-water seeping through/ the biological and other soil-development processes in the thicker or thinner layer of soil covering the surface.

It should be mentioned that in the interior of the lithoclase network the water seeping into the network of cracks in the limestone from the surface humous soil layers no longer has the possibility to come into contact with gases whose compositions are significantly different qualitatively or quantitatively from that of the soil atmosphere. Thus, the carbonic acid values obtained in the soil are retained by the water almost unchanged as it seeps through right to the well ventilated cavern, or again emerges at the surface, where the new environment causes the evaporation of the dissolved carbon dioxide brought by the water to give a new equilibrium corresponding of course to lower p_{CO_2} values.

The temperature of the water seeping down under the action of gravitation in the lithoclase system in the interior of the maturing karstic limestone mass is practically unchanged during its passage as a rule. It is known that the vertical systems of crevices in which the water trickles down are at the same time the means for thermal exchange by convection; these can ensure the uniform temperature corresponding to the yearly average for the region almost perfectly in the entire system above the karst water level. In this way, the thermal factor can have an effect on the CO₂ content of the water more or less only in the upper soil zone of the infiltration, where its action is exerted in accordance with the values given in Table 3.

It must be noted, however, that compared to the effect of the partial pressure differences the role of the soil temperature differences which occur in nature in practice and which quantitatively effect the CO₂ dissolution is much smaller, even in extreme cases; at any p CO₂ level whatever, water at 0 °C is able to dissolve only about twice as much as that dissolved by water at 20 °C. Thus, in the karst dynamic evaluation of the role of the thermal factor, in agreement with the majority of the present climatic karstmorphogenetic authors /H. LEHMANN, TROMBE, BÖGLI, WISSMANN, GVOZDETKII, etc./, the correct procedure is probably to conceive the effect of the temperature levels characterizing the individual climate zones of the earth not in the direct SCHLOESING sense, but as an indirect effect composed of the increase of the material transport per unit time /intensity of dissolution/, soil atmospheres containing more CO₂, and more intense soil processes assisted by the higher thermal level.

For example, the warm rain in the tropics /because of its higher temperature/ is able to absorb less CO₂ gas than the cold water from melted snow in the polar regions for instance. This greatly overrides and subordinates the fact that at the same time the dissolution process is faster in the tropics because of the higher temperature, while the production of carbon dioxide is much more vigorous as a result of the intense inorganic and biogenetic soil processes. As the final complex resultant here, therefore, the possibilities of the high-level saturation of the water with CO₂ will nevertheless be much more favourable.

We have now arrived at the possibility of taking up a well-founded clear stand in the international controversy which arose following the ominous attitude of CORBEL /1959/: it is not CORBEL, but H. LEHMANN and the climatic geomorphologists following in his footsteps who are right. These latter not only recognise temperature and precipitation-order differences in the climate factor, but with a real DOKU-CHAYEV-ian attitude can take into account each of the geomorphological, pedological, biological, hydrological, chemical, etc. factors resulting in the climatic zonality, and also their simultaneous complexity even /JAKUCS 1970/.

We have so far studied the roles of the partial gas pressure and the temperature from among the factors determining the capacity of the water to absorb carbon dioxide from the atmosphere or the soil-air. It still remains, therefore, to analyse the pressure and time factors.

It is not necessary to provide a special proof that the ability of water to absorb CO₂ increases with increasing pressure. It is sufficient to point to the clear example of soda-water, which provides a striking experimental

confirmation. It is less well known, however, that among the natural actualities of karsts the possibility of the development of the pressure factor is ensured by the periodicity of the rainfall, water or soil atmosphere penetrating from the karst soils into the lithoclasts of the limestone, depending on whether there is an ample surface water supply or a lack of it.

When there is a very rainy period the cracks in the limestone fill with water. However, if these filaments of water running down under the action of gravity can no longer obtain a continuous replacement during a drier period, then their continuity is interrupted; a weak vacuum forms and air from the soil is sucked into the network of cracks.

When it again rains, of course, the "entrances" above the lithoclast network soil again become covered with water, and this now forces the gas mixture already in the cracks to greater depths.

The cracks are frequently so thin that the laws relating to adhesion and capillary attraction overcome the forces striving to arrange the system according to density. Thus, alternating gas and liquid phases migrate downwards, somewhat similarly to what can be seen in a thermometer with a broken mercury thread.

Even in this stage certain hydrostatic pressure modifications will affect the equilibrium of the water - gas system, since if there would be no pressure distortions arising from the gravitational and adhesive forces in the system of cracks, then the water there would not move downwards. These pressure changes, however, are still not so

significant that they would substantially modify the water - carbon dioxide phase-equilibrium formed in the soil. Nevertheless, the water sooner or later reaches a level where there is no longer a free, empty route towards the depths. This is either because the sub-layer is impermeable and so there is no network of cracks, or because the existing network is already filled with water. The arriving water therefore finds its path blocked, begins to build up, and together with the accompanying air bubbles is forced to seek an outflow possibility in the lateral directions.

This is not easy, because the cracks, which now should lead away the water in the three-dimensional network to the karst level in the horizontal directions, that is superficially or linearly, are themselves no more developed or wide at the beginning of the karst maturation than the vertical lithoclase system. The water units /and also the gas bubbles between them/ are therefore compelled to enter zones where the hydrostatic pressure increases strongly. The pressure here will be higher, the greater the water supply from the surface and the lower the possibility for lateral movement. Thus, a karst water zone flowing at high pressure develops in the interior of the karsts /JAKUCS 1960, 1968a, 1968b/, in which the water will become enriched by further amounts of dissolved carbon dioxide, since the pressure is also exerted on the gas mixture, and this is now practically compressed into the solution.

In this zone, where the spaces and cracks in the rock which ensure the passage of water are now filled with water, a pressure of several atmospheres may result /mining observations!/. The pressure is particularly considerable in the

lower level of the karst water zone where the movement is horizontal. Experience has shown that it is not rare, especially in the case of young karsts, for the pressure to exceed 10 atmospheres even.

We shall return to this question later, but in connection with the magnitude and effectiveness of the role of the pressure it must be noted here that this comes about essentially via a certain modification of equation /3/. According to the classical gas law of BOYLE-MARIOTTE, the volume of a given mass of gas measured at various pressures is inversely proportional to the pressure. The factor 1.9634 given in equation /3/ is therefore not a constant value, but is valid only for the case of a pressure of one atmosphere. For a pressure of 2 atmospheres this factor must be doubled, for a pressure of 10 atmospheres it must be multiplied by 10, and so on, since the weight of the gas too has increased in these proportions as a result of the compression.

Let us now assume a case where the value of p_{CO_2} does not change, that is the composition of the gas mixture remains constant. Then, in the gas mixture - water system at $10^\circ C$ and a pressure of 5 atmospheres with $p_{CO_2} = 0.002$, the total amount of CO_2 gas absorbed by the water will increase from the value 4.69 mg/l given in Table 3 to a value of $1.194 \cdot 0.002 \cdot 5 \cdot 1.9634 = 23.45$ mg/l. Since this number is exactly five times the amount of CO_2 dissolved from the air at a pressure of 1 atmosphere, it might be said that up to a certain limit the capacity of water to dissolve CO_2 varies with the pressure in accordance with the product of the capacity at 1 atmosphere and the number of atmospheres /HENRY's law/.

At the same time, it is essential to bear in mind a very important factor, which can modify the validity of the above hypothesis considerably under karstic conditions. This is the fact that as a result of the effect of pressure the composition of the original gas space changes in the closed two-component system in the interior of the karst, in so far as the greater the pressure, the less the partial pressure of CO₂ in the gas space. This is related with the fact that the absorption coefficients /L/ of different gases are not the same. Those of the gases playing the most important parts in the gas space are substantially smaller than that of CO₂ /for oxygen at 20 °C L = 0.031; for nitrogen at 20 °C L = 0.015; and for carbon dioxide at 20 °C L = 0.0878/.

This has the result that with the increase of pressure the carbon dioxide is relatively quickly exhausted from the heterogeneously composed gas system, and the higher pressure values are practically powerless to increase still further the total CO₂ content of the solution. Because of this it is our opinion that as regards the increase of the karst water it is the earlier pressure increases /e.g. from 1 to 10 atmospheres/ which play the essential role, the unit pressure increases for hydrostatic pressures of up to even 100 atmospheres on the deep karst waters having much lower effects on the direct increase of the dissolution.

As a result of the differences in the absorption coefficients for the three gases, it is understandable that the composition of the original gas mixture is finally very changed, and at extremely high pressures it consists almost exclusively of the most difficultly soluble nitrogen. /This is why nitrogen always predominates in the gas bubbles coming to the surface with the waters of karst springs rising from great depths, for example at Miskolc-Tapolca, in the warm springs in Buda, etc./

It may explicitly be stated, therefore, that the fact of the extensive increase of the dissolved CO₂ as a result of the increase of the hydrostatic pressure is very important from the point of view of the development of the karsts, because the state of the karst corrosion level /the embryonic cavern formation level/ in the depths is indicated by the practically secondary aggressivity of the water.

It is plausible that with the later release of the pressure, when the water reaches a wider cavity or surface where its directions of movement are determined by the slope, the gases compressed into it in the pressure zone will come out of solution; this is also accompanied by the cessation of the secondary hydrocarbonate dissolving capacity.

In such places, however, the water is as a rule not only released from the hydrostatic load, but at the same time also comes into an environment where the $p \text{ CO}_2$ is much lower than in the infiltration soil zone. Thus, further CO_2 will leave the solution in accordance with the difference in $p \text{ CO}_2$ values for the two air spaces, until a new equilibrium has been established corresponding to the $p \text{ CO}_2$ /and temperature/ of the new air space as governed by equation /3/.

In the literature one can find the evaluation of the role of pressure from a different viewpoint. Thus, of the Hungarian authors particularly DUDICH /1932/, CHOINOKY /1940/ and KESSLER /1938/ long ago stressed the importance of the changes in the surface tension of the water. It is a matter here of the tensile force occurring as a result of the surface tension on the outer surface of the water drop moving from the crevice in the rock to the ceiling of the cavern,

and also of the surface pressure decreases affecting the karst waters flowing on the convex slope and reduced to spray in the waterfalls. These factors simplify and accelerate the diffusion of the excess CO_2 into the gas-space.

With this we have already come into contact with the fourth determinant of the CO_2 absorption, the time factor. This is similarly a very important factor in the establishment of equilibrium between the dissolved CO_2 and that in the gas-space. That is, equilibrium is established in accordance with equation /3/ in a time process.

The time requirement for the establishment of equilibrium in a given case could be expressed by an artificial formula, but even today it cannot be correctly calculated in practice. The time requirement for the establishment of the equilibrium could be determined only as the complex resultant of many factors which themselves are also variable, but in several cases the precise mathematical formulation of even these partial factors is not possible.

In the following only some of the determinants which have the most important effects on the time factor will be mentioned, simply to illustrate what has already been said.

The greater the contact surface between the solution and the gas-space, the faster the process. But the reaction is also accelerated by the increase of the temperature. The reaction time is likewise shortened if the gas-space is in turbulent motion. If there is also such a turbulent mixing motion in the solution space, this further increases the reaction rate. The qualitatively and quantitatively different absorption surfaces in the reaction space /e.g. mineral and organic soil particles/ either promote or hinder the absorption, depending on their specific properties.

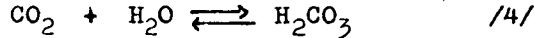
The rate of the dissolution process will be different if the water is not in contact with limestone, than when there is such a contact and hence the formation of hydrocarbonate can begin simultaneously with the gas absorption. Many more such modifying factors could be listed.

It is undoubtedly true that this is one of the most complicated problems in the interpretation of the corrosion. Nevertheless, in spite of its complexity, at the same time it is one of the key questions in climatic karst morphogenetic analysis, and one of the fundamental modifying factors of the dynamism of limestone dissolution.

So far, the dissolution of the gaseous CO_2 has been studied only as regards the properties of the liquid and the gas phases, and the solid phase /in this case the limestone/ has not yet been considered. In the following we again return to the analysis of the combined steps of the dissolution process when all three phases are present simultaneously.

It is understandable that with the introduction of this new factor the equilibrium relations and interactions become even more extensive than when the binary system of only the water and the CO_2 gas was examined. The question is even more complex, because here it is no longer a matter only of simple dissolution or dissociation, but of the chemical transformations which are closely connected with the dissociation and other physical dissolution equilibria.

According to literature data /HOLLUTA 1927, PIA 1953/, at 4 °C 0.7 % of the absorbed carbon dioxide content of the solution is converted to carbonic acid, in agreement with the well-known classical equation:



If it comes into contact with CaCO_3 , the carbonic acid reacts with it, with the formation of calcium hydrocarbonate, and hence the limestone dissolves:



Naturally, the carbonic acid bound to the calcium and thus practically used up is replaced from the physically absorbed CO_2 of the solution /until this is exhausted/, so that in practice almost the entire amount of dissolved CO_2 takes part in the process of dissolving the limestone.

According to the calculations of TILIMANS and HEUBLEIN /TILIMANS 1932, 1940/, the lower the concentration of carbonic acid, the more effective is equation /5/ in the direction of the upper arrow. Thus, in weak solutions /e.g. a total dissolved CO_2 concentration of 10-15 mg/l the reaction proceeds almost completely according to the upper arrow, while in more concentrated solutions it does so only up to a certain fraction of the total absorbed CO_2 . The reason for this is that in the case of higher ionic concentrations accessory or equilibrium free carbonic acid must be present in amounts

gradually increasing with the concentration, so as to maintain the dissolved hydrocarbonate in equilibrium.

Figure 2 shows the most important bonding and equilibrium interactions governing the dissolution of $\text{Ca}/\text{HCO}_3/2$.

For every mg of CO_2 dissolved in the water, the solution can dissolve 2.2723 mg CaCO_3 , independently of the temperature of the system. Accordingly, therefore, at low concentrations, where it is still not practical to consider the equilibrium free carbonic acid requirements, from a knowledge of the amount of dissolved CO_2 the limestone dissolving capacity of the solution can be calculated at once. For example, the concentration of dissolved CO_2 in water saturated at 10°C with air of normal composition $/p \text{CO}_2 = 0.0003/$ is 0.70 mg/l. Without the entry of further gaseous CO_2 into the solution, this is capable of dissolving $2,2723 \cdot 0.7 = 1.59$ mg of limestone as hydrocarbonate.

Naturally, even under laboratory conditions such a limestone-containing solution can not occur in the presence of the solid phase, since the solid phase, since the solution is also bound to contain about 13 mg/l calcium carbonate as a result of the primary carbonate dissolution in accordance with equation /1/. It must also be pointed out that in addition to carbon dioxide oxygen too is absorbed in natural soil waters. This is used in bacterial and inorganic oxidation processes in the soil, and hence in the majority of cases gives rise to an increased CO_2 content /TUCAN 1933, CHRAMUSEV 1941, VENKOVITS 1949, STEFANOVITS 1956/.

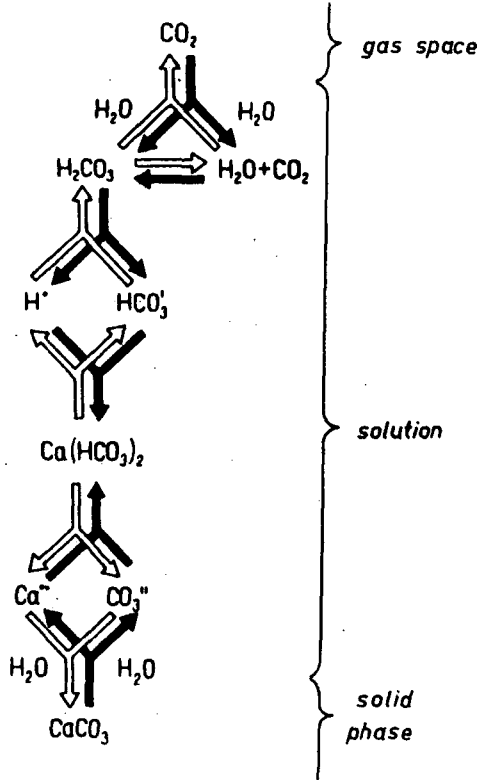


Figure 2. The most important chemical conditional components of the hydrocarbonate dissolution of limestone. The equilibrium shifts according to the black arrows and to the white arrows result in dissolution and precipitation of limestone, respectively. At the same time the equilibrium shift for the system can occur either only according to the black arrows, or only according to the white ones.

At higher concentrations the question of the dissolution is more involved only in so far as the total amount of carbon dioxide in the water can not be taken into account in the hydrocarbonate reaction, because the equilibrium free carbonic acid requirement must also be considered. It should be noted, however, that the amount of this accessory free carbonic acid now depends not only on the quantity of dissolved limestone, but also on the temperature of the environment.

In the presence of CaCO_3 , therefore, the absorbed carbon dioxide is to be found simultaneously in the following forms in the aqueous solution:

1. In the form of hydrocarbonate bound to the calcium.
2. In the form of accessory, or equilibrium free carbonic acid necessary to maintain the hydrocarbonate in solution.
3. In the form of aggressive carbonic acid, in amounts possibly exceeding the equilibrium requirements of the hydrocarbonate. This latter free carbonic acid obtained its name from its corrosion capacity, since not being involved in the existing solution equilibria, it can facilitate the dissolution of further limestone.

If the relative amounts of these three functional types of carbonic acid in a solution are known, it can be decided whether the given solution is capable of further corrosion, whether the limestone and the carbonic acid are in equilibrium, or whether the solution is supersa-

turated and hence its tendency is to deposit calcium carbonate.

In the international chemical and hydrological literature, it was earlier possible to find the equilibrium free carbonic acid indexes only for waters at 17 °C /TILIMANS 1932, LAPTYEV 1939, TROMBE 1951, 1952, BÖGLI 1960, etc./. For this reason PAPP /1954, 1956/ calculated the accessory free carbonic acid contents of waters in limestone - carbonic acid equilibrium for all the more important concentrations and temperatures. The values most likely to be required in karst research work are given in Table 4.

It is also possible to decide from the Table the extents of the function of the temperature changes controlling the solubility, in so far as water possessing an unchanged total CO₂ content is capable of retaining less calcium carbonate in solution at higher temperatures, because at higher temperatures the amount of equilibrium free CO₂ necessary to keep the calcium hydrocarbonate in solution increases. If, for example, the total CO₂ content of a water in limestone - carbonic acid equilibrium at 5 °C is 957 mg/l /330 mg/l of this is bound CO₂, and 627 mg/l accessory carbonic acid/, this will keep in solution Ca/HCO₃/₂ corresponding to 749.7 mg/l solid CaCO₃. If the temperature of the water is now raised to 14 °C, however, with an unchanged total amount of carbonic acid, the limestone-dissolving capacity decreases, and thus the solution becomes supersaturated. In this case the 957 mg/l total CO₂ content now assumes a different distribution: the bound CO₂ will be 306.2 mg/l, and the equilibrium CO₂ 650.8 mg/l, and accordingly the quantity of CaCO₃ which can be kept in solution by the water cannot be more than 969.1 mg/l. This means that 749.7 - 696.1 = 53.6 mg CaCO₃ must separate out from each litre.

Table 4

The bound and accessory CO₂-requirements of the hydrocarbonate dissolution of the limestone

CaCO ₃ mg/l	bound CO ₂	accessory CO ₂ /mg/l/ at the following temperatures							
		5 °C	7 °C	9 °C	10 °C	11 °C	12 °C	15 °C	20 °C
8,92	3,85	-	-	-	-	-	-	-	-
17,85	7,70	-	-	-	-	-	0,01	0,01	0,01
26,77	11,5	0,03	0,03	0,04	0,04	0,04	0,04	0,04	0,05
35,70	15,6	0,07	0,07	0,07	0,08	0,08	0,08	0,09	0,00
44,52	19,4	0,14	0,15	0,16	0,17	0,17	0,18	0,19	0,23
53,55	23,5	0,23	0,24	0,26	0,26	0,27	0,28	0,30	0,35
62,47	27,3	0,38	0,41	0,44	0,45	0,46	0,48	0,51	0,59
71,40	31,5	0,54	0,58	0,61	0,63	0,65	0,67	0,73	0,84
80,32	35,3	0,79	0,84	0,89	0,92	0,95	0,97	1,06	1,23
89,25	39,2	1,06	1,11	1,18	1,21	1,25	1,28	1,40	1,62
98,17	43,0	1,39	1,52	1,61	1,65	1,71	1,75	1,91	2,21
107,1	47,1	1,82	1,93	2,04	2,10	2,17	2,23	2,43	2,81
116,0	50,9	2,27	2,50	2,65	2,79	2,81	2,89	3,15	3,64
124,8	55,0	2,90	3,08	3,26	3,36	3,45	3,56	3,88	4,48
133,8	58,8	3,60	3,82	4,04	4,16	4,28	4,41	4,81	5,56
142,8	62,7	4,30	4,56	4,83	4,97	5,12	5,27	5,57	6,64
151,7	66,5	5,22	5,53	5,86	6,03	6,21	6,40	7,48	8,06
160,6	70,6	6,14	6,51	6,90	7,10	7,31	7,53	8,21	9,49
169,6	74,4	7,29	7,73	8,19	8,43	8,70	8,96	9,75	11,2
178,5	78,5	8,45	8,96	9,49	9,77	10,1	10,4	11,3	13,1
187,4	82,5	9,92	10,4	11,0	11,3	11,7	12,1	13,2	15,2
196,3	86,5	11,4	11,9	12,7	13,0	13,4	13,8	15,1	17,4
205,3	90,4	13,0	13,6	14,5	14,9	15,3	15,8	17,3	19,9
214,2	94,2	14,6	15,4	16,4	16,8	17,3	17,8	19,5	22,5
223,1	98,1	16,6	17,5	18,6	19,1	19,7	20,2	22,1	25,6

CaCO ₃ mg/l	bound CO ₂	accessory CO ₂ /mg/l/ at the following temperatures							
		5 °C	7 °C	9 °C	10 °C	11 °C	12 °C	15 °C	20 °C
232,0	102,1	18,6	19,7	20,8	21,5	22,1	22,7	24,6	28,7
241,0	106,0	20,9	22,1	23,4	24,1	24,8	25,5	27,9	32,3
249,9	110,0	23,2	24,6	26,1	26,8	27,6	28,4	31,0	35,9
258,8	113,9	25,9	27,4	29,1	29,9	30,3	31,7	34,6	40,0
267,7	117,9	28,6	30,3	32,1	33,0	34,0	35,0	38,2	44,2
276,7	121,8	31,6	33,4	35,4	36,5	37,5	38,6	42,2	48,8
285,6	125,6	34,6	36,6	38,8	40,0	41,1	42,3	46,2	53,4
294,5	129,5	38,0	40,3	42,7	44,0	45,2	46,5	50,3	58,7
303,3	133,5	41,5	44,0	46,6	48,0	49,4	50,8	55,5	64,1
312,3	137,4	45,4	48,1	51,0	52,5	54,0	55,6	60,7	70,1
321,3	141,4	49,3	52,3	55,4	57,0	58,7	60,4	65,9	76,2
330,2	145,3	53,5	56,8	60,1	61,9	63,7	65,6	71,6	82,8
339,1	149,2	57,8	61,3	64,9	66,8	68,8	70,8	77,3	89,4
348,0	153,1	62,6	66,4	70,3	72,4	74,5	76,7	83,7	96,9
357,0	157,1	67,5	71,5	75,8	78,0	80,3	82,7	90,2	104,3
365,9	161,0	72,9	77,2	81,9	84,3	86,7	89,3	97,5	112,7
374,8	165,0	78,4	83,0	88,0	90,6	92,3	96,0	104,8	121,1
383,8	168,9	84,3	89,2	94,6	97,4	100,2	103,2	112,6	130,2
392,7	173,5	90,2	95,5	101,3	104,3	107,3	110,5	120,5	139,3
401,6	176,9	96,5	102,2	108,3	112,5	114,8	118,2	128,9	149,0
410,5	180,7	102,8	108,9	115,4	118,8	122,3	125,9	137,4	158,8
419,5	184,6	109,8	116,3	123,3	126,9	130,7	134,5	146,8	169,7
428,4	188,5	116,8	123,8	131,2	135,1	139,1	143,2	156,2	180,6
437,3	192,4	124,5	131,9	139,8	143,9	148,2	152,5	166,8	192,4
446,2	196,4	132,2	140,1	148,4	152,8	157,3	161,9	176,6	204,2
455,2	200,3	140,3	148,7	157,6	162,2	167,0	171,9	187,5	216,8
464,1	204,2	188,5	157,4	166,8	171,7	176,8	182,0	198,5	229,5
473,0	208,1	159,0	166,9	166,8	182,0	187,4	192,9	210,5	243,3

CaCO ₃ mg/l	bound CO ₂	accessory CO ₂ /mg/l/ at the following temperatures							
		5 °C	7 °C	9 °C	10 °C	11 °C	12 °C	15 °C	20 °C
481,9	122,1	166,5	176,4	186,9	192,4	198,1	203,9	222,5	257,2
490,8	216,0	176,1	186,6	197,7	203,6	209,6	215,7	235,4	272,1
499,8	220,0	185,7	196,8	208,6	214,8	221,1	227,6	248,3	287,0
508,7	223,9	196,1	207,8	220,2	226,7	233,4	240,3	262,1	303,0
517,6	227,9	206,5	218,8	231,9	238,7	245,8	253,0	276,0	319,1
526,5	231,8	217,3	229,8	244,0	251,2	258,6	266,2	290,4	335,8
535,5	235,7	228,1	241,8	256,2	263,8	271,5	279,5	304,9	352,5
553,4	243,5	251,9	266,9	282,9	291,2	299,8	307,9	336,6	389,2
571,2	251,2	276,5	293,0	310,5	319,6	329,1	338,0	369,5	427,2
589,0	259,2	303,3	321,5	340,7	350,7	361,0	370,8	405,4	468,7
606,9	267,1	331,9	351,8	372,8	383,8	395,1	405,8	443,6	512,9
624,7	275,0	362,8	384,5	407,5	419,5	431,5	443,5	484,9	560,6
642,6	282,7	394,1	417,6	442,5	455,6	469,0	481,7	526,6	608,8
660,3	290,6	428,1	453,6	480,7	494,9	509,5	523,3	572,1	661,4
678,3	298,5	464,0	491,7	521,1	536,5	552,3	567,2	620,1	716,9
696,1	306,2	500,8	530,7	562,5	579,0	596,1	602,2	669,3	773,8
714,0	314,1	540,6	572,9	607,1	625,0	643,4	660,9	722,5	835,3
731,8	322,0	582,5	617,3	654,1	673,4	693,3	712,0	778,5	900,0
749,7	330,0	627,0	664,4	704,1	724,8	746,2	766,4	837,9	968,7
767,5	337,7	671,8	711,9	754,5	776,7	799,6	821,2	897,8	1038
785,4	345,6	720,1	763,1	808,7	832,9	857,1	880,3	962,4	1112
803,2	353,6	770,5	816,6	865,3	890,8	917,1	941,9	1029	1190
821,1	361,2	822,0	871,1	923,2	950,3	978,4	1004	1098	1271
838,8	369,1	877,1	929,5	985,1	1014	1044	1072	1172	1355
856,8	377,0	934,7	990,5	1049	1080	1012	1142	1249	1444
874,6	385,0	995,6	1055	1118	1151	1185	1217	1330	1538
892,5	392,7	1056	1119	1186	1221	1257	1291	1411	1632

The above relations can be illustrated in the form of a complex graph /Fig. 3/ derived from the values given in Table 4. The abscissa of this diagram shows the total amount of CO_2 in the solution, and the ordinate the amount of CaCO_3 dissolved, both in mg/l. If the previously determined total CO_2 content or dissolved CaCO_3 content is projected into the system, the chemical nature of the water studied can readily be read off.

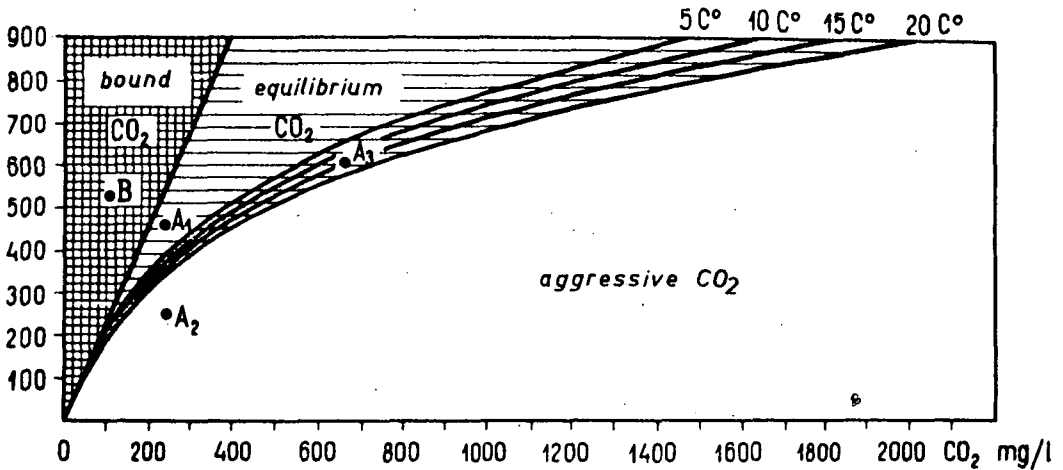


Figure 3. Amounts of bound carbonic acid and equilibrium free carbonic acid in waters in limestone - carbonic acid equilibrium at 5, 10, 15 and 20 °C.

If, for instance, the total amount of CO_2 is 250 mg/l, and the amount of dissolved CaCO_3 is 450 mg/l /in this case the point A_1 denotes the state of the water on the diagram/, it can be seen at once that the water is strongly supersaturated, and thus CaCO_3 will precipitate from it. If, on the other hand, the amount of dissolved CaCO_3 is only 250 mg/l, then the water is strongly aggressive, because this state /position A_2 on the diagram/ is in a region of unsaturation. At the same time, another water, defined on the basis of analysis by point A_3 , and obviously at equilibrium at 12 °C, becomes aggressive if cooled to a temperature lower than this, and supersaturated at higher temperatures.

Of course, it is not possible to find a water sample which can be denoted by a point inside the region of bound CO_2 /e.g. B/.

It must be mentioned here that the aggressivity of a water does not lead to the immediate unfailing dissolution of limestone in all cases, for if it is possible for the carbon dioxide to evaporate out of the solution, then this in an alternative method for equilibrium corresponding to the new conditions to be established.

The route by which the system strives to attain equilibrium in a given case depends on many involved factors. In a closed system where there is no possibility for the evaporation of CO_2 into the gas-space, the equilibrium can only be restored by the increase of the corrosion. /This is why waters circulating in a closed network of tubes are so corrosion-dangerous, even if they are only slightly aggressive./

The aggressivity of water can readily be calculated numerically from a knowledge of the total CO_2 content and the carbonate hardness* from Table 4.

By the combined use of Table 3 and 4, still further very significant considerations are possible as regards karst research. Thus, from a knowledge of the $p \text{CO}_2$ values of the soil atmosphere determining the chemical nature of the water in the infiltration zone, the extent of limestone transportation can be calculated.

Let us now consider a concrete example of the use of these Tables for such calculations:

The water has been acted on in the soil by air with a partial pressure of 8/100, and the water - carbonic acid equilibrium has been established by contact with a gas-space of this concentration. The temperature of the soil during the infiltration period was 5 °C. In this case the water has a total carbonic acid content of 224 mg/l /see Table 3/. The aggressive water dissolves up limestone in its network of crevices, and during the seepage downwards a limestone

* The carbonate hardness of the water is determined by the sum of the calcium and magnesium hydrocarbonates dissolved in it. One German hardness degree means 10 mg/l CaO, or 17.85 mg/l CaCO_3 . Thus, for example, water with a German hardness degree of 22, if the hardness is caused only by calcium salts, contains $17.85 \cdot 22 = 392.7$ mg/l of CaCO_3 dissolved as hydrocarbonate.

- carbonic acid equilibrium is established. The amount of dissolved limestone is found from Table 4; this is done by seeing how the total CO_2 content of 224 mg/l is divided at 5 °C between bound and equilibrium free carbonic acid. Values of these are sought such that their sum corresponds as closely as possible to 224 mg/l. In this case the sum of 157.1 mg/l bound carbonic acid and the related 67.5 mg/l accessory CO_2 approximates well to the desired number, the difference being negligible. It only remains to read off the value of CaCO_3 in mg/l from the first column of Table 4, giving in the present case a figure of 357.0 mg/l.

From this, however, we have so far found only the amount of limestone which may be dissolved by the karst water under favourable conditions. As to how much may be reprecipitated from the solution under given conditions, such as those in a cavern, this can be learnt only from further calculations.

If, for instance, the karst water in question finds its way into a cavity in a cavern at 10 °C where the partial pressure of CO_2 in the air is 2/100, then if a suitable amount of time is available a new equilibrium state will develop corresponding to the CO_2 content of the new atmosphere; this means, according to Table 3, that the total CO_2 content of the solution will be 46.9 mg/l /the rest evaporates/. Next, by the already known method, the CaCO_3 value must be found from Table 4 corresponding to

$$\text{/bound CO}_2\text{/ + /equilibrium CO}_2\text{ at 10 °C/ = 46.9 mg/l}$$

Concretely, it is seen that under the new equilibrium conditions the solution contains only about 100 mg/l CaCO_3 . The difference in the two values $357 - 100 = 257$ mg/l/ is thus the desired amount which is transported by the water from one point of the karst to another.

It must be noted here, however, that although a good approximation may often be obtained, nevertheless these calculations can not always be regarded as entirely realistic. The problem is that in nature there are several factors which cannot always be calculated at present, e.g. the time factor, the other gases dissolved in the water /mainly oxygen/, the modifying actions of the humus and root acids in the zone of decomposition of the organic matter, etc. These vary from case to case, and at times may make such a significant contribution that they can change the quantity of limestone dissolved considerably.

From another aspect, it also emerges from Table 4 that while the amounts of CO_2 bound as hydrocarbonate increase linearly with the amount of dissolved CaCO_3 , the increase of the equilibrium free carbonic acid is of a progressive nature; that is, the successively increased amount of dissolved CaCO_3 can be kept in an equilibrium state only by the ensurance of continually more rapidly increasing amounts of accessory free carbonic acid. The relation is expressed by the following formula, due to TILIMAS:

$$\text{equilibrium CO}_2 = \frac{\sqrt{\text{bound CO}_2^3}}{K_t} \quad /6/$$

where the factor K_t is a constant for a given temperature.

It follows clearly from all this that during the mixing of equilibrium hydrocarbonate solutions of different concentrations /e.g. natural soil and karst waters/ carbon dioxide will become surplus in the solution, that is the solution will become aggressive. Depending on the conditions, this carbonic acid excess will evaporate from the solution, or may lead to the dissolution of more CaCO_3 . In such a way, a secondary dissolution process can come about; this is known as mixing corrosion. IAPTYEV /1939/ was the first to draw attention to its existence, but its role in karstification has been suitably stressed only by BÖGLI /1963/, and in Hungary by ERNST /1964/ and BALÁZS /1966/.

The extent of the reaggresivity appearing on mixing the solutions is the greater, the larger the difference in the initial hydrocarbonate concentrations of the mixed equilibrium solution components. Only a minor role is played in this question by the temperature factor.

Let us now confirm the above findings by the examination of some concrete data from Table 4.

Before mixing, the equilibrium solutions A and B had the following compositions:

solution A /9 °C/	CaCO ₃ content dissolved as hydrocarbonate	151.7 mg/l
	bound CO ₂ content	66.5 mg/l
	equilibrium free CO ₂ content	5.8 mg/l

solution B /9. °C/	CaCO ₃ content dissolved as hydrocarbonate	508.7 mg/l
	bound CO ₂ content	223.9 mg/l
	equilibrium free CO ₂ content	220.2 mg/l

If it is assumed that /for the sake of simplicity/ equal volumes of A and B are mixed, then the product, solution C, will have the following composition:

solution C /9 °C/	CaCO ₃ content dissolved as hydrocarbonate	$= \frac{151.7 + 508.7}{2} = 330.2 \text{ mg/l}$
	bound CO ₂ content	$= \frac{66.5 + 223.9}{2} = 145.2 \text{ mg/l}$
	free CO ₂ content	$= \frac{5.8 + 220.2}{2} = 113.0 \text{ mg/l}$

If the adequate carbonic acid values related to the dissolved CaCO₃ content of solution C, 330.2 mg/l, are looked for in Table 4, it is seen that the value for the bound carbonic acid is in agreement, whereas at 9 °C the requirement of solution C for accessory free carbonic acid is equivalent to only 60.1 mg/l CO₂. Thus, as a result of the mixing there will be a 113.0 - 60.1 = 62.9 mg/l excess of in statu nascendi CO₂ in solution C, which means that the solution now contains aggressive carbonic acid.

Using the same method as in the above example, we have calculated the concentrations of the aggressive carbonic acid adequate CO_2 released on mixing equal quantities of waters with carbonate hardnesses of 0-50, at 10 °C. The results are shown in Table 5.

Table 5

0°	0	5°																	
5°	0,5	0	10°																
10°	3,7	1,3	0	15°															
15°	12,4	7,3	2,4	0	20°														
20°	29,2	20,6	10,8	3,1	0	25°													
25°	57,4	43,9	28,8	14,9	2,7	0	30°												
30°	98,8	80,0	58,8	35,8	18,1	4,7	0	35°											
35°	157	132	102	73,5	45,2	22,3	6,0	0	40°										
40°	234	200	164	126	87,6	64,1	24,8	5,8	0	45°									
45°	333	293	247	198	149	102	61,4	30,3	7,7	0	50°								
50°	448	407	351	292	230	169	117	70,1	32,4	7,8	0								

Amounts of aggressive carbonic acid CO_2 /mg/l/ formed on the 1:1 mixing of equilibrium waters at 10 °C with German hardness factors of 0-50

The same data are also plotted graphically in Figure 4. Hence, by means of the continuity of the lines on the diagram the amounts of in statu nascendi CO₂ relating to the intermediate values can be read off directly.

It has been mentioned that temperature changes act in a subordinate way, and only as a modifying factor, in the quantitative control of the aggressive carbonic acid formed on mixing. To illustrate this, let the temperature of solution A in the above example now be 9 °C still, but that of solution B 15 °C. After mixing, the temperature of solution C will then be 12 °C.

In accordance with the new temperature, the equilibrium amount of CO₂ in solution B is now 262.1 mg/l compared with the previous 220.2 mg/l. The other characteristics of solutions A and B do not change. Thus, the total quantity of free CO₂ in solution C can be calculated as:

$$\frac{5.86 + 262.1}{2} = 133.9 \text{ mg/l}$$

Since the equilibrium carbonic acid requirement of solution C at 12 °C is 65.6 mg/l, then as a result of the mixing there will be a free CO₂ excess of 133.9 - 65.6 = 68.3 mg/l. This is only 5.4 mg/l more than the amount released on mixing the solutions which were both at 9 °C.

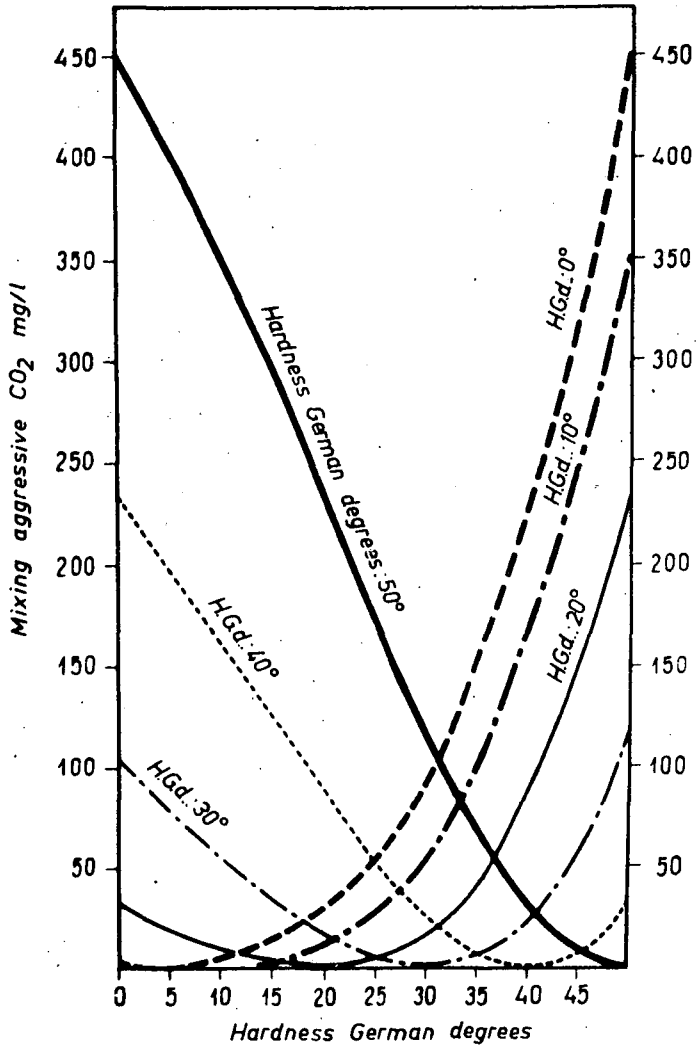


Figure 4. Amounts of aggressive carbonic acid /CO₂ mg/l/ formed in the 1:1 mixing of karst waters of various hardnesses at 10 °C.

It can be stated that the larger the temperature difference between the two solutions to be mixed, the greater the amount of aggressive CO₂ formed. For example, if the temperature of solution A is only 5 °C, while that of solution B is 15 °C, it can readily be calculated that 60.7 mg/l aggressive CO₂ will be produced in the solution C formed with a temperature of 10 °C. This is 7.8 mg/l more than the amount formed on mixing solutions A and B, both at 9 °C.

The diagrams of Figure 5 were plotted from the results of calculations in this connection. They show the amounts of in statu nascendi CO₂ released on the mixing of equal volumes of soft water not containing dissolved chemicals /pure rain-water/ and equilibrium karst waters of various carbonate hardnesses. The calculations were performed for systems at 5, 10 and 20 °C.

Of course, only part of the unbound carbonic acid released during the mixing processes is able to dissolve further CaCO₃, for the remainder is required to keep the newly dissolved amount in equilibrium.

The very important question from the point of view of the karst researcher, as to how much limestone can be dissolved by secondary dissolution as a result of the aggressive CO₂ released when solutions mix, can be answered most simply in the following way with the use of Table 4:

The concentrations of all the bound and accessory CO₂ in the equilibrium solutions A and B are added together, and the result divided by two. The value so obtained is the total amount of CO₂ in one litre of solution C. The bound CO₂ values and the accessory CO₂ values at the

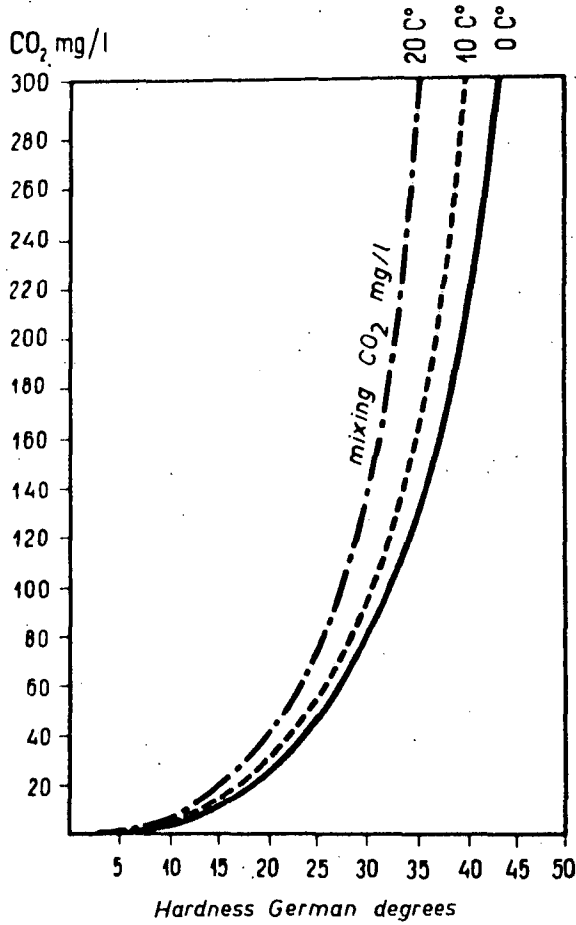


Figure 5. Amounts of in statu nascendi CO₂ /mg/l/ released on the 1:1 mixing of soft water not containing dissolved chemicals and equilibrium waters of various carbonate hardnesses in systems at 5, 10 and 20 °C.

appropriate temperature are taken from the columns of the Table and added together. This is done until the sum is the same as the already calculated total CO₂ concentration. The sought-for amount of secondarily dissolved limestone is given by the difference between the CaCO₃ value relating to the row giving the coincidence, and the average CaCO₃ value for solution C.

Let us now illustrate the above on the example of solutions A and B at 9 °C.

The total CO₂ supply of solution C, obtained by summing the concentrations in solutions A and B, and averaging, is:

$$\frac{66.5 + 5.8 + 223.9 + 220.2}{2} = 258.2 \text{ mg/l.}$$

The average CaCO₃ supply in this same solution C has already been calculated: 330.2 mg/l.

If the sums of the bound CO₂ value and the accessory CO₂ value at 9 °C, as obtained from the relevant columns of the Table, are compared with the total CO₂ content of 258.2 mg/l, with a little interpolation it can readily be found that, in an equilibrium state, this corresponds to 166.7 mg/l bound CO₂ and 91.5 mg/l accessory CO₂.

Next, the amount of CaCO₃ actually dissolved by solution C can easily be calculated from the formula N.V, where N = the concentration of bound CO₂ in the solution /mg/l/, and V = 2.2723, a constant.

In this case, therefore, in its new equilibrium state solution C will be capable of retaining in solution an amount of hydrocarbonate equivalent to $166.8 \cdot 2.2723 = 378.8$ mg/l CaCO_3 . Thus, by means of the mixing corrosion the solution is capable of the secondary dissolution of $378.8 - 330.2 = 48.0$ mg/l limestone.

Contradictory views can be found in the literature as regards the role of the mixing corrosion in karstification. According to BÖGLI, one of the leading authorities in this question, /BÖGLI 1963a, 1963b/, it is important only in those zones deep in the karst where the passages are completely filled with water and there are no well-ventilated air-spaces into which the released aggressive carbonic acid might evaporate. In contrast with this, it has been pointed out by Hungarian research workers among others /GÁNTI 1957, MAUCHA 1960, CZÁJLIK 1962, ERNST 1964/, and supported by numerous observations and analyses, that the recent corrosion of cave waters appearing to be strongly supersaturated may frequently be observed even in wide passages.

It is clear that the debate can be settled by the study of the rates of the dissolution and gas-diffusion processes, and in this respect we consider that the opinions of FRANKE /1965/, ERNST /1964/ and BALÁZS /1966/ most closely reflect the truth. They hold the view that mixing karst waters of different hardnesses may become capable of dissolution even when in contact with a free air-space. The evaporation of carbon dioxide from the solution takes place only at the surface of contact of the two media, and is therefore a slow process, whereas the rate of mixing of the water components is much higher since they are in turbulent motion, and so the water molecules in contact with the rock surfaces are continually being replaced by fresh ones.

It is correctly stressed by BALÁZS /op. cit. p. 184/ that: "The greatest concentration differences in the waters of our karst caves may appear when the very hard karst waters originating from infiltration and stagnating in the depths, are mixed with the alluvial waters from melting snow or sudden summer storms." In his view, on each such occasion the intrushing waters may permit the secondary dissolution of several hundred kilograms of limestone in the larger Hungarian caves, quite apart from their possible inherent aggressive effects.

As regards the many details of the question and the extent of the dissolution under natural conditions, of course, a final answer can be expected only after the collection and evaluation of many more careful observations. Nevertheless, it can already be stated with certainty that the mixing corrosion, as one of the important components of the hydrocarbonate dissolution process, plays a really factorial role in the development of the karsts.

We have so far made a multilateral analysis of the multicomponent regularities of the interactions governing the amount of carbon dioxide in the atmosphere, the carbonic acid content of the water, and the amount of limestone dissolved by the solution. It still remains to introduce a very important equilibrium relation which must be assumed between the partial pressure of CO_2 in the atmosphere /or soil atmosphere/ and the solution in limestone - carbonic acid equilibrium. This is at the same time the

most refined and the most time-requiring equilibrium of the hydrocarbonate dissolution. This is why in many cases under natural conditions the karst water does not attain equilibrium.

Here, therefore, the amount of dissolved carbonic acid consumed during the hydrocarbonate formation process will show up as a loss in the equilibrium balance for the first-degree binary system atmosphere - water, and this results in a renewed tendency for the dissolution of atmospheric carbon dioxide. Thus, in other words it could be said that since the amounts of carbonic acid bound to calcium /or magnesium/ as hydrocarbonate are no longer involved in the absorption equilibrium, and only the accessory CO_2 is in an equilibrium relation with it /according to equation /6//, then the accessory carbonic acid in the solution must come into a new adequate equilibrium relation with the partial pressure of CO_2 in the air /according to equation /3//. That is, as shown by Figure 2 the progress of the limestone-dissolving process produces effects all the way back to the atmosphere in contact with the solution, and leads to the absorption of further carbon dioxide from this.

In this way, the water in contact with a limestone will continue to absorb carbon dioxide from the atmosphere until equilibrium according to equation /3/ has been established between the atmospheric CO_2 and the accessory /equilibrium/ free carbonic acid of the solution. This does not mean, therefore, that the SCHLOESING Table /Table 3/ is unusable, but simply leads to the

limitation that in the case of a dissolved limestone content the values of the Table must be referred to the equilibrium CO₂ values of Table 4 at the appropriate temperature.

Let us now consider the above in the case of an earlier used example.

In an equilibrium state at 5 °C, the partial pressure of CO₂ in ordinary atmospheric air, 0.0003, corresponds to 0.84 mg/l dissolved CO₂. We have already seen that if this carbonic acid value is regarded as stationary, then the hydrocarbonate-dissolving capacity of saturated water in air is very slight. If, however, in accordance with the extent of its exhaustion in the hydrocarbonate dissolution, the means and the time are available for further absorption from the atmosphere, by the time this dissolved CO₂ value of 0.84 mg/l is again established as the accessory carbonic acid in equilibrium with the hydrocarbonate, the water has already dissolved up about 81.97 mg/l CaCO₃ /calculated by interpolation from Table 4/.

Thus, it becomes quite understandable that even the softest karst waters, which are typical of completely bare karsts and alpine, snow-zone karsts, may contain as much as 80-100 mg dissolved limestone per litre. /BALÁZS /1963 reports a value of only 104 mg/l for the limestone content of the Cadisha spring which emerges from the bare karstic mass of the 3000 m high Kornet es Saouda in the Lebanon. The Lodowe Zródło karst spring which surfaces in the Polish West Tatra contains 75-85 mg/l CaCO₃. KESSLER determined

80-100 mg/l CaCO_3 in karst waters originating from the depths of bare Albanian karsts at a plateau height of 2000 m . BAUER /1964/ and TELL /1961/ mention karst waters of similarly low concentrations in the Austrian Alps and in Sweden, respectively./

As regards order, the above values agree quite well with the results of theoretical calculations obtained by other means /MILLER 1952, MANDY 1954, GÁNTI 1957, ERNST 1961, MARKÓ 1961, FRANKE 1967/.

The question of the time requirements of the series of hydrocarbonate limestone dissolving reactions/ has been dealt with recently by BÖGLI /1960, 1963/. He found that as regards the time requirements of the hydrocarbonate dissolution of limestone it is convenient to divide the dissolution into four stages as follows.

In the first stage, which is very rapid, the simple physical dissolution of the CaCO_3 occurs as in equation /1/. The partial conversion of the CO_2 absorbed by the water to give carbonic acid, and the dissociation of this to H^+ and HCO_3^- ions has already occurred earlier. The time needed for the first stage is about 1 sec.

The second stage, during which the carbonate ions / CO_3^{2-} / of the limestone become bonded to the H^+ ions of the carbonic acid, can not be distinguished in time from the first. Because of the loss of CO_3^{2-} ions, however, the equilibrium which develops in the first stage is disrupted in this one, and this means that further physical /carbonate/ dissolution of the limestone must of necessity follow.

In the second stage, however, the equilibrium of the physically and chemically dissolved CO_2 in the water is also disturbed. Because of this the third stage soon begins. This is the transformation of the physically dissolved carbon dioxide to carbonic acid.

BÖGLI found the time requirements of this third stage to be about 1 minute. During this time the original CO_2 supply of the water is used for the hydrocarbonate dissolution of the limestone. According to him the fourth stage then begins: further amounts of carbon dioxide are absorbed from the atmosphere. By this means the series of dissolution reactions continues until the final equilibrium state has developed. This is defined by an equilibrium of the highest order via the repeated transfer of the amounts of limestone dissolved as hydrocarbonate and the carbon dioxide content of the atmosphere in contact with the water.

In this way, the dynamics of the dissolution of the limestone in the fourth stage are determined by the rate of absorption of gaseous carbon dioxide, and this is very low. According to BÖGLI, 24-60 hours is required for the establishment of the final equilibrium. FREAR and JOHNSTON, on the other hand, mention several days.

A very important factor involved in the determination of the absorption and diffusion rates is the temperature. When this is low it substantially slows down the gas absorption, but when it is high the absorption can be accelerated many times /because of the more rapid motion of the molecules colliding on the surface between the

media, and hence because of their higher number/. According to the generally accepted view of FEITKNECHT, for every temperature increase in the solution of 10 °C the rate of reaction is approximately doubled. The duration of the absorption and /for the motion of the process in the opposite sense/ diffusion is also significantly affected by the size of the surface per unit amount of water. Accordingly, the greater the exchange surface /water drops, aerosol, waterfall, etc./, the greater the material transport in a given unit time. The extent of the material transport in unit time, however, is also influenced considerably by the hydrostatic pressure, as was shown above.

The modern analytical results of GERSTENHAUER and PFEFFER /1966/ are in contradiction with the theory of BÖGLI regarding the duration of the dissolution; in particular they cast strong doubt /and rightly so in our view/ on the reaction rates of BÖGLI's stages 1, 2 and 3. Thus, if we accept BÖGLI's breakdown of the series of reactions into four stages, because of the didactic viewpoints of its descriptiveness, we must know that in practice the individual stages never follow one after the other. They always come about simultaneously and mutually, that is they are not separated into discrete stages. The reversibility chain of the series of dissolution reactions, with the narrowest permeable capacity during unit time, in fact takes place according to the uppermost arrow shown in Figure 2, that is between the water and the atmosphere. But in our view for just this reason it is more correct to formulate the problem so that the rate of hydrocarbonate dissolution in the three-
- component system is determined fundamentally from the

beginning on by this absorption or diffusion material transport capacity. The factorial components and comprehensive regulating conditions of this, however, have already been studied in detail earlier.

III. Non-karstic corrosion of limestone

While the simple physical /carbonate/ and carbonic acid /hydrocarbonate/ processes of dissolution of limestone are collectively termed karstic corrosion, the corrosive denudative processes connected with the other chemical effects form the non-karstic group of the chemical weathering of limestone.

In this respect very many, mainly double decomposition chemical reactions may be given, primarily consisting of the mutual reactions of the limestone and the chemically active constituents of the soils of various compositions. These are in part predominantly the effects of the chemical products formed by means of inorganic mineralization weathering processes in the soil; in part the organic and inorganic chemical products /mainly various acids/ of plant and animal refuse decomposing under aerobic or anaerobic conditions in the humus levels of the soils, and selected by soil microorganisms and the roots of plants living in the soils. On the action of all such soil processes which from our point of view can be generally regarded as non-karstic corrosive factors, calcium from the limestone gives up its carbonate bonding and a completely new calcium compound is formed. This is either transported away

in the form of its solution, or it remains for a longer or shorter time in an autochthonous state as a solid reaction product.

The up-to-date pedological, agrochemical and soil-biological handbooks /FEKETE 1952, 1958, STEFANOVITS 1959, 1963, FEHÉR 1954, PEJVE 1961, DI GLÉRIA 1962, BALLENEGGER and DI GLÉRIA 1963, FEKETE, HARGITAI and ZSOLDOS 1964, BECK 1968/ give a detailed analysis of humification processes whose dynamics are greatly affected by the climatic conditions, and which are characteristic of soil-levels and of soils of various compositions. Of the dissimilation compounds formed during the aerobic decomposition of organic substances in the soil, or as a result of the life-functions of the microorganisms in the soil, those which are the most important from the point of view of the corrosion of the limestone are formic acid, oxalic acid, acetic acid, propionic acid, lactic acid, the various root secretions, and of course the already discussed carbonic acid. These all form the corresponding calcium salt with the limestone.

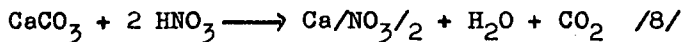
It would still be difficult today to decide the extents of the separate individual dynamics of these compounds in the dissolution of limestone, all the more so because their effects in the soil are complex and simultaneous, and because the proportion of their regional contribution in the soil solution is not constant either. Their overall effect, however, is more easy to establish. It is sufficient for the demonstration of this for us to point to one of the most classical experiments of biochemistry. In this a seed was

caused to germinate on a marble surface; during even a short time the minute root of the seed left a clearly visible trace on the smooth surface of the CaCO_3 /SACHS 1865/.

The most important of the corrosive organic acids formed during the aerobic decomposition processes /under anaerobic conditions too, but much more slowly/ are the fulvic and crenic acids. In an acidic soil medium /under aerobic conditions/ these exert particularly active dissolving effects, and their resulting salts remain in solution.

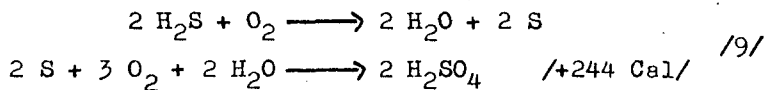
Limestone is also actively corroded by the humous and humic acids of the soil, but in acidic medium the calcium salts of these are precipitated.

The effects of inorganic acids and salts formed in the course of biogenetic or mineral weathering processes in the soil are also of great importance from the point of view of the corrosion denudation of the limestone. Of these, sulphuric and nitric acids among others, as very strong acids, even in very low concentrations give rise to the fact that their double decompositions with limestone are unidirectional /irreversible/:



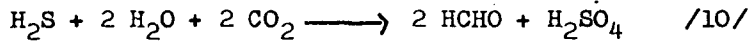
It should be noted that the formation of the above acids in the soil may be connected with various processes. Thus, the sulphuric acid most frequently originates from the inorganic or biogenic /activity of sulphur bacteria/ oxidation of sulphides /e.g. pyrite/ and hydrogen sulphide /H₂S/. The starting compounds are formed from the chemical weathering of inorganic constituents of the soil, and more frequently from the mainly anaerobic process of decomposition of the organic refuse derived from living organisms. FEHÉR /1954/ reports that, for example, it is typical of the functioning of Achromatium, Beggiatoa, Thiotrix, Thioploca, and other genera belonging to the colourless sulphur bacteria, that they first split off sulphur from the H₂S released in the decomposition of proteins. This is stored in the form of drops in their plasma, and gradually consumed so that energy is obtained for their C-assimilation.

In outline, the process is as follows:



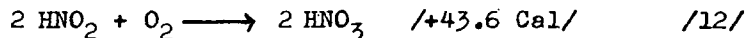
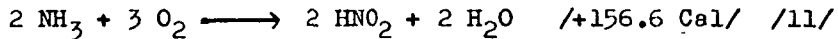
According to FEHÉR /op. cit. p. 115/, the life-function of the hydrogen sulphide-oxidizing purple bacteria proceeds via a process which is somewhat more complex than the previous one. These are also aerobic organisms, and thus they oxidize hydrogen sulphide and assimilate chemosynthetically. With the intervention of bacteriochlorophyll and bacterioerythrin in their plasma, they are also able to use the photoenergy of the sun, and so they carry out photosynthetic assimilation at the same time

too. With the help of the hydrogen sulphide, and with the use of photoenergy, the CO_2 is deoxygenated and hydrogenated to give formaldehyde:



Of the sulphur bacteria living in the soil, Thiocystis, Thiospirillum, Rhabdomonas, Rhodotheca, and other genera can be classified in this group.

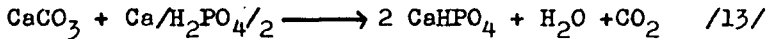
Ammonia, resulting likewise mainly from organic decomposition, must be regarded as the most probable starting substance for the formation of nitric acid in the soil. With the intervention of bacteria, the ammonia is oxidized in the soil to acids which have a corrosive action on limestone. It was already demonstrated in the last century by VINOGRADSKII /1892/ that the process of nitrification in the soil is carried out by two metabiotically cooperating groups of bacteria: the nitrite-forming and the nitrate-forming bacteria. The nitrite bacteria oxidize the ammonia to nitrous acid, and the nitrate-forming bacteria oxidize the nitrous acid so formed to nitric acid:



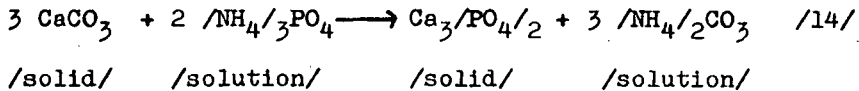
Of course, neither the nitrous acid nor the nitric acid can remain in the soil in an unbound state, and immediately on their formation react with the cations of the soil and the limestone to form their nitrites or nitrates /equation /8//.

It should be noted that a certain amount of nitric acid can enter the soil from the atmosphere. As stressed by FINDEISEN /1939/, HARRASOWITZ /1954/, KILINSKI /1958/ and REITER /1960/, the summer rain almost always contains nitric acid. This is formed in the atmosphere by the electric discharges accompanying the storms. Although it is not yet possible to determine the extent of the participation in the corrosion by the acid of such an origin because of the unsolved difficulties of the experimental methods, nevertheless it must be assumed that this amount of acid constitutes a considerable denudative factor, especially in the tropics, where the atmospheric electric discharges are of consistent frequency.

Because of its importance among the possibilities of non-karstic corrosive processes on the limestone by the soil solutions, the action of phosphate solutions must definitely be stressed. For example, that of calcium dihydrogen phosphate, which reacts with limestone to form calcium monohydrogen phosphate:

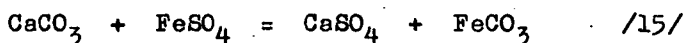


Ammonium phosphate, which occurs in the rendzina soils of limestone mainly as a product of the decomposition of animal bodies, acts on the limestone as follows:



If the volume of the solid tricalcium phosphat formed /calculated from the molecular volume/ is taken as 100, then the unit volume of the mass of the starting CaCO_3 solid is 110.7. That is, the double decomposition reaction is accompanied by a volume decrease, and so it can occur even deep in the structure of the limestone. If the new solid mass formed, which replaces the original solid mass in the stone, does not occupy the same volume as the starting mineral, then unfilled spaces remain around the new formation in the structure of the stone /its pore volume increases/. Thus, the aggressive solution can penetrate deep into the interior of the layers, where the total transformation of the original material can be brought about in a very large thickness.

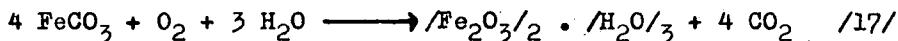
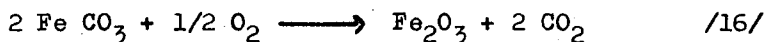
This refers too to all such double decompositions which result in solid reaction products and in which the product formed has a higher mineral density than that of the initial solid phase. This is also the case for the reactions of iron salts with limestone, e.g.:



Naturally, not only the sulphates, but also the salts of iron formed with other strong acid radicals can easily bring about the double decomposition, all the more so because the molecular volume of iron carbonate /siderite/ is smaller /30.6/ than that of calcium carbonate /36.9/, and the exchange takes place in a 1:1 molecular ratio.

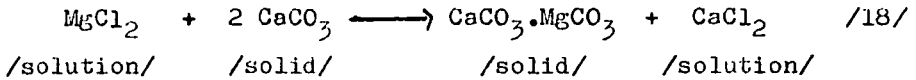
In the above example, a classic case was introduced of the epigenization of limestone by iron salts.

It is worth mentioning that the solid siderite crust forming on the boundary surface between the limestone and the soil can not impede the effect of the iron sulphate in the direction of deeper limestone levels, because it is soon dissolved by the carbonic acid-containing soil solutions. In the presence of water and oxygen /the karstic rendzina soil reaction medium corresponding to this/, the ferrous carbonate is sooner or later oxidized to ferric oxide. This can accumulate in the soil as either the dark-red haematite $/\text{Fe}_2\text{O}_3/$, or the brown limonite $/2\text{Fe}_2\text{O}_3 \cdot 3\text{H}_2\text{O}/$:



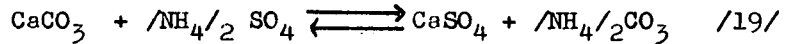
It is probable that, similarly to the process of siderite formation, many other exchanges occur for the limestone. Thus, the formation of glauconite and flintstone, and the substitution of sulphides of the heavy metals /e.g. pyrite/ in place of the calcite crystals /pseudomorphs/, may be the results of equivalent processes. ELIE DE BEAUMONT /referred to in CAYEAUX /1935// is surely right when he assumes that the explanation of the subsequent dolomitization of limestone in contact with sea-water is also to be sought in this phenomenon. In this case the magnesium chloride of the sea-water reacts with the calcium carbonate and the well-known double salt dolomite is obtained. The reaction is possible even in the interior of the limestone, because the density of the double carbonate, which is more poorly

soluble than the calcium and magnesium carbonates, is much higher than that of calcite.



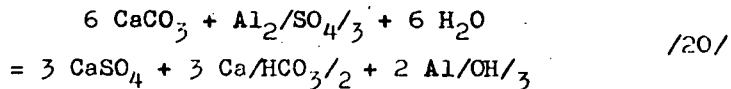
$$2 V_1 = 73.8 > V_2 = 63.6$$

A role is also played in the non-karstic corrosion of limestone by the often appreciable amounts of ammonium sulphate in the soil:

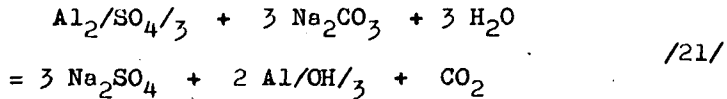


It should be taken into account, however, that this reaction takes place in the direction of the upper arrow only when the compounds on the right side of the equation are more dilute than those on the left side. That is, the continuity of transportation away of the calcium sulphate and ammonium carbonate formed must be provided by some effect such as the motion of the soil water, or a biogenic force. If, however, for some reason the concentrations are higher on the right side /e. g. in sea-water/, the double decomposition proceeds in accordance with the lower arrow /BECK 1968/.

The corrosive action of aluminium sulphate, if present in the soil, proceeds in an interesting way:



It should be noted that the above reaction takes place only in neutral or acidic soils. In aqueous medium the acidically hydrolyzing aluminium sulphate is first consumed in decreasing the alkalinity of the soil, according to the following equation:



In order for us to form as complete a picture as possible of the effective forms of non-karstic corrosion, which has received fairly scant attention in the karst literature, it must still be mentioned that the dynamism of the dissolution may be significantly affected not only by the possible formation of chemical bonds with the various ions occurring in the soil solutions, but also by their simple mass actions, and in such a way that the effectiveness of the hydrocarbonate dissolution is indirectly regulated merely by their presence.

It is known that the solubility of a salt decreases in the presence of a salt of the same ion, whereas it is increased by the presence of a salt of different ions /GRÓH 1939, PAPP 1954, etc./. Thus, for example, the solubility of sodium chloride can be decreased by magnesium chloride, but at the same time the alkali chloride content /NaCl, KCl/ of the water increases the solubilities of calcium carbonate, calcium sulphate, calcium phosphate, etc. The solubility of CaCO_3 is similarly increased if the soil solution contains magnesium chloride, for example.

The modern view is that the hydrocarbonate dissolution of limestone is decreased by the other dissolved carbonates by changing the value of the denominator, K_t , of equation /6/ to a minimum. This has the result of increasing the accessory carbonic acid necessary to keep the calcium in solution. Because of this, natural karst waters, which always contain greater or lower amounts of carbonates of other ions too, generally require more free carbonic acid than that calculated on the basis of the calcium hardness.

According to PAPP /op. cit. p. 39/, the practical extent of this increased requirement is for the most part not appreciable, because especially in hard waters, which usually contain other carbonates too, compared to the otherwise considerable amounts of equilibrium carbonic acid the extra CO_2 requirement is insignificant. At the same time, it should also be taken into account that the decrease in the solubility of CaCO_3 produced by salts of the same ion may easily be counterbalanced by those salts of other ions which are always present in various amounts in karst waters, and which generally tend to increase the solubility of the CaCO_3 . This means, therefore, that in the majority of cases the data of Table 4 calculated from equation /6/ do in fact correspond to the accuracy requirements of a good approximation.

In this short section we have by no means exhausted all the factorial possibilities of non-karstic corrosion of limestone in nature. We have attempted to stress only the most important of the components, as reflected in their observable, objective effects.

It must be pointed out, however, that the extremely fact-requiring and labour-consuming study of the many details and the evaluation of the extents of the effects involved in the question forms one of the perspective tasks of karst science. The future answers to these questions may provide very much to the entire karst science; it may well be that a certain fundamental reevaluation of the present axioms of karst theory will be necessary. But primarily perhaps we can still expect an impetus from the growth of the information available, in the settling of divergent problems of the denudation-dynemism of the climate zonal karst regions.

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